

Assessment of mercury contamination of the Svitava and Svatka rivers and muscle of chub (*Leuciscus cephalus* L.) in the urban agglomeration of Brno in the Czech Republic

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Abstract

This study focused on the total mercury and methylmercury content in the muscle of chub (*Leuciscus cephalus* L.), the total mercury in the river sediments and the evaluation of health risks associated with fish contamination. Chub were caught at seven localities on the Svatka and Svitava rivers in the agglomeration of Brno in 2008. The results were compared to those obtained from the same sites in 2007. Total mercury was determined by atomic absorption spectrophotometry using an AMA 245 analyzer, and methylmercury was determined by gas chromatography (using an electron-capture detector) after acid digestion and toluene extraction in chub muscle. The highest concentrations of total mercury and methylmercury (0.12 ± 0.14 and 0.07 ± 0.02 mg·kg⁻¹ fresh weight, respectively) were found in Svatka before junction (south of Brno), whereas the lowest concentration of mercury and methylmercury in chub (0.06 ± 0.01 and 0.04 ± 0.01 mg·kg⁻¹) was detected in Svitava before junction with the Svatka River. Total mercury in sediments ranged from 0.01 to 1.05 mg·kg⁻¹ dry weight, the highest value was detected in the sediment from Rajhradice. The lowest content (0.01 mg·kg⁻¹) was at Kníničky. Hazard indices calculated for the selected localities showed no health risk for either a standard consumer or a fishing family. Fish from the Svitava and Svatka rivers show very low mercury concentration and hazard index and their consumption poses no health risk from total mercury and methylmercury contamination.

Czech river, Leuciscus cephalus, methylmercury, sediment, total mercury

Mercury pollution is a ubiquitous problem with atmospheric deposition contaminating watersheds in areas far from anthropogenic or natural atmospheric point sources (Swain et al. 1992). Several forms of mercury are present in the aquatic environment, including elemental, ionic, and organic (Morel et al. 1998). In most freshwater, ionic mercury in the divalent state (Hg^{II}) is the predominant form, whereas in most fish species more than 95% of mercury occurs as methylmercury (MeHg). Thus, the conversion of ionic mercury to MeHg is an important link in the bioaccumulation of mercury in fish and ultimately in its toxicity to humans and wildlife (Eisler 2006). Methylmercury is the most toxic form and bioaccumulates in fish primarily through dietary uptake (WHO 1990). The level of bioaccumulation is a function of age, species, and trophic transfer. Methylmercury exposure can affect growth, reproduction, development and survival in fish (Weiner and Spry 1996). Halbach (1995) reported that bioaccumulation of mercury in fish and its toxicity in humans can be attributed to the high affinity of mercury for sulphur-containing proteins such as metallothioneins and glutathione. The highest concentration of mercury in fish is contained in muscle, and thus muscle is a good indicator of mercury contamination (Čelechovská et al. 2007).

A project of the Ministry of Education Youth and Sport of the Czech Republic comprised a two year study (2007 and 2008) focusing on mercury contamination of the Svitava and Svatka rivers (Czech Republic). In addition to mercury contamination, the project included assessment of persistent organic pollutants using selected

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biochemical markers in chub (Blahová et al. 2009, 2010), and passive sampling for monitoring endocrine disruptors (Grabic et al. 2010). The present report summarizes mercury contamination in chub in 2008, and compares the results with those from 2007 (Kružíková et al. 2009).

The main aims of the present study were to determine total mercury (THg) and MeHg (methylmercury) concentrations in chub muscle from 7 localities on the Svitava and the Svatka rivers in the Brno agglomeration, to assess THg concentration in sediment at these localities, to evaluate health risks associated with consumption of fish from these sites and to compare data from 2007 and 2008.

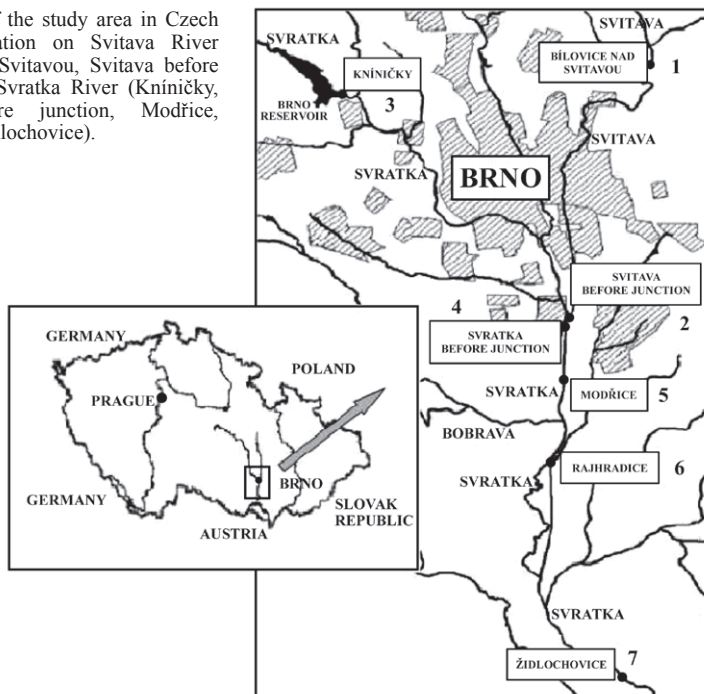
Materials and Methods

Sampling sites

Mercury contamination was assessed in the Rivers Svitava and Svatka which run through Brno, the second largest city in the Czech Republic (population 366,680). Brno is an important industrial city with highly developed engineering, chemical, textile, and food-processing industries. Domestic waste, sewage, and other effluents from industrial sources are the most likely source of persistent pollutants in the area's aquatic ecosystem.

The Svatka River runs 29 km through the city and is the major source of water for the Kníničky reservoir, a popular recreation area in the northwest part of Brno. The Svitava River flows through the city for about 13 km and merges into the Svatka River downstream of Brno and is the main tributary to the Svatka River. Seven sites were chosen to evaluate the influence of Brno on mercury contamination of fish and sediments. Two sites were on the Svitava River: (1) Bílovice nad Svitavou, 18 river km; and (2) Svitava before junction, 0.6 river km. Five sites were on the Svatka River: (3) Kníničky, 56.2 river km; (4) Svatka before junction, 40.9 river km; (5) Modřice, 38.7 river km; (6) Rajhradice, 35.0 river km; and (7) Židlochovice, 30.0 river km. Sites 1 and 3 situated upstream of Brno characterize conditions upper of the city. Sites 2 and 4 are situated downstream of Brno and above the confluence of the two rivers. Site 5 is located downstream of a sewage treatment plant and characterizes the Svatka River in areas of waste water effluent. Sites 6 and 7 situated below Brno represent the cumulative effects of all other sites. Weirs are situated between sites 1 and 2, 3 and 4, 5 and 6 and 6 and 7. Fig. 1 shows the locations of the sites.

Fig. 1. Map of the study area in Czech Republic: location on Svitava River (Bílovice nad Svitavou, Svitava before junction) and Svatka River (Kníničky, Svatka before junction, Modřice, Rajhradice, Židlochovice).



Sampling of fish and sediments

In 2008, 119 chub (*Leuciscus cephalus* L.) were captured by electrofishing, immediately weighed, measured for total length, and scales collected for age determination. Sex was determined macroscopically. For THg and MeHg analysis, muscle samples were taken from the cranial area dorsal to the lateral line, placed in polyethylene bags, labelled, and stored at -18 °C. Table 1 shows the main biometric characteristics of fish collected in 2008.

Sediment was sampled from the same locations in February, March, and September. At each location (both years), composite bottom sediment was collected into dark glass bottles, lyophilized in the laboratory, and stored at -18 °C. Sampling was validated in accordance with ISO 5667 12 norm.

Table 1. Biometric values of sampled chub (*Leuciscus cephalus* L.)

Sample site	n	Total length (cm)	Weight (g)	Age (years)
1 Bilovice nad Svitavou*	15	25.2 ± 3.3	182 ± 97.5	4 ± 1.3
2 Svitava before junction*	17	23.9 ± 4.5	241 ± 107.3	3 ± 0.9
3 Kníničky**	14	25.1 ± 7.5	242 ± 101.2	3 ± 1.1
4 Svatka before junction**	17	29.4 ± 4.9	311 ± 179.0	4 ± 0.9
5 Modřice**	24	26.1 ± 4.8	365 ± 239.0	3 ± 1.1
6 Rajhradice**	17	29.2 ± 4.7	310 ± 179.5	3 ± 0.8
7 Židlochovice**	15	26.9 ± 3.3	231 ± 89.2	2 ± 0.7

*the Svitava River; **the Svatka River

Determination of THg and MeHg

The THg content in muscle and in sediments was determined by cold vapour atomic absorption spectrometry using AMA 254 analyzer (Altec Ltd., Czech Republic).

The MeHg content was determined in the form of methylmercury chloride by gas chromatography (Caricchia et al. 1997). Samples were prepared by acid digestion and extraction with toluene (Maršálek and Svobodová 2006). A Shimadzu capillary gas chromatograph with an electron captured detector GC 2010A (Shimadzu Kyoto, Japan) was used for analysis. A capillary column DB 608 (30 m × 0.53 mm × 0.83 µm; J&W Scientific Chromservis, Czech Republic) was used. Data evaluation was made with GC Solution software (Shimadzu Kyoto, Japan).

Limits of detection for THg and MeHg were 1 µg·kg⁻¹ and 21 µg·kg⁻¹, respectively. The limit of detection was set as a sum triple the standard deviation of a blank and a blank mean value. The accuracy for THg and MeHg values was validated using standard reference material BCR-CRM 464 (Tuna Fish, IRMM, Belgium).

The total mercury and MeHg concentrations in fish muscle are given in mg kg⁻¹ fresh weight (FW), and THg in sediment is given as dry weight (DW).

Health hazard assessment

The hazard index was calculated according to Kannan et al. (1998) using a reference dose for THg (0.3 µg·kg⁻¹body weight day⁻¹) set by the United States Environmental protection Agency (US EPA). To determine the maximum safe consumption of fish, a provisional tolerable weekly intake limit (PTWI) of 1.6 µg MeHg kg⁻¹ body weight week⁻¹ was used (WHO 1990). The amount of fish safe to be eaten per week was calculated.

Statistical analysis

The statistical analysis of the data was performed using STATISTICA 8.0 for Windows (StatSoft CR). The data were analyzed with the parametric ANOVA Tukey's HSD test.

Results

Total mercury and methylmercury in fish

The content of THg and MeHg was measured in chub, an omnivorous fish suitable for the monitoring of aquatic ecosystems. The sampled fish were of similar age. The main biometric data for sampled chub are shown in Table 1. No significant differences in the age of the fish were found. Since no significant sex-related differences in THg and MeHg content were found, data for both sexes were combined.

Mercury and methylmercury content in the muscle of indicator fish in 2008 is given in Fig. 2. No significant differences in the mercury content (THg and MeHg) were found in fish from the various sites. The lowest value of THg and MeHg was found at site 2 (0.06 ± 0.016 and 0.037 ± 0.011 mg·kg⁻¹, respectively) and at site 5 (0.06 ± 0.083 and 0.04 ± 0.039

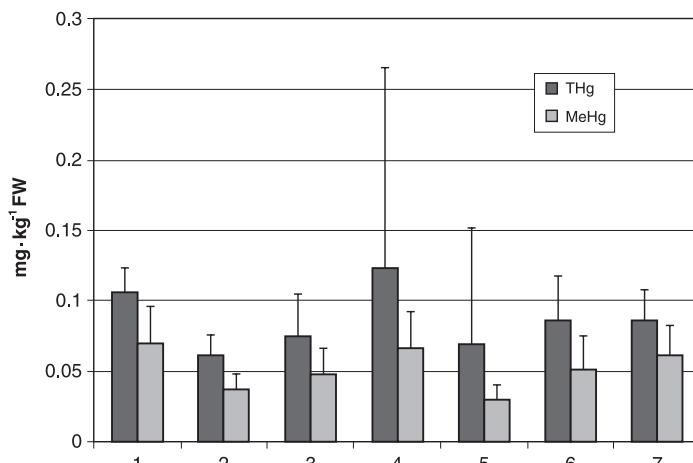


Fig. 2. Total mercury and methylmercury content in the muscle of chub (*Leuciscus cephalus* L.) in 2008

mg·kg⁻¹, respectively). The highest THg and MeHg concentrations were detected on the Svatka River (site 4) (0.12 ± 0.14 and 0.06 ± 0.016 mg·kg⁻¹, respectively).

Total mercury in sediment

Sediments for THg determination were sampled 3 times during 2008 (Table 3). The characteristics of the sediments were not consistent from site to site. As in 2007, sediments from site 1 and 3 were arenaceous while other sediments were sloughy.

In 2008, the highest concentration of THg was found on site 6 and reached 1.05 ± 0.13 mg·kg⁻¹ DW. In 2007 the highest concentration of THg in the sediment was found at site 4 (1.21 ± 0.24 mg·kg⁻¹). Sites with the lowest amounts were similar in both years (site 3 and 1).

Discussion

Total mercury and methylmercury in fish

Kružíková et al. (2009) reported mercury content in fish sampled in 2007 at the same sites. In 2007 significant differences among sites were found, in contrast to results in 2008. A reduction of the mercury content was observed at all sites compared to 2007 with the exception of site 4. A large decrease in THg and MeHg was found at site 6 (2.25 times) from 0.18 to 0.08 mg·kg⁻¹ THg. This was not found in the mercury content of the sediment from this site. Although the mercury concentration in 2007 did not exceed the hygienic

Table 2. Average total mercury content (mg·kg⁻¹) and % methylmercury to total mercury in chub (*Leuciscus cephalus* L.) in 2007 and 2008

Sample site	Total Hg		Methyl Hg to total Hg (%)	
	2007***	2008	2007***	2008
1 Bílovice nad Svitavou*	0.13 ± 0.06	0.11 ± 0.019	55.8 ± 25.1	70.5 ± 25.9
2 Svitava before junction*	0.11 ± 0.03	0.06 ± 0.016	68.2 ± 21.6	61.6 ± 19.1
3 Kníničky**	0.11 ± 0.04	0.07 ± 0.030	75.7 ± 26.3	66.4 ± 18.3
4 Svatka before junction **	0.10 ± 0.03	0.12 ± 0.143	85.6 ± 18.7	55.1 ± 30.2
5 Modřice**	0.08 ± 0.02	0.06 ± 0.083	52.8 ± 40.2	61.9 ± 24.3
6 Rajhradice**	0.18 ± 0.08	0.08 ± 0.032	87.8 ± 14.4	59.5 ± 19.5
7 Židlochovice**	0.13 ± 0.02	0.08 ± 0.022	74.2 ± 18.8	70.6 ± 16.5

* Svitava River; ** Svatka River, *** Results from Kružíková et al. (2009)

Table 3. Total mercury levels (mg·kg⁻¹ dry weight) in sediment in 2008

Sample site	2007***	2008			Mean ± SD
	Mean± SD	February	March	September	
1 Bilovice nad Svitavou*	0.11 ± 0.03	0.031	0.061	0.046	0.05 ± 0.01
2 Svitava before the junction*	0.59 ± 0.23	0.190	0.205	0.905	0.43 ± 0.41
3 Kníničky**	0.06 ± 0.01	0.017	0.013	0.010	0.01 ± 0.01
4 Svatka before the junction**	1.21 ± 0.21	0.491	0.781	0.239	0.50 ± 0.27
5 Modřice**	0.71 ± 0.18	0.426	0.636	0.637	0.57 ± 0.12
6 Rajhradice**	1.15 ± 0.18	0.993	0.950	1.194	1.05 ± 0.13
7 Židlochovice**	0.51 ± 0.19	0.859	0.766	0.318	0.65 ± 0.29

* The Svitava River; ** the Svatka River, ***results from Kružiková et al. (2009)

limit, it is a felicitous finding that mercury contamination was reduced. The THg content was below the hygiene limits set by Commission Regulation (EC) No. 1881/2006 in either 2007 or 2008. The limit for total mercury in fish products and muscle meat of non-predator fish is 0.5 mg kg⁻¹ and 1.0 mg kg⁻¹ for selected fish species.

Many studies have examined the levels of mercury contamination in Czech rivers such as the Elbe (Dušek et al. 2005; Žlábek et al. 2005; Maršálek et al. 2006; Maršálek and Svobodová 2006), the Ohře, the Vltava, the Morava, and the Sázava (Kružiková et al. 2008a). In general, it has been demonstrated that mercury contamination of Czech rivers is low and does not exceed the limit of 0.5 mg·kg⁻¹ FW, which is in agreement with the present study.

The proportion of methylmercury to total mercury ranged from 55.1 to 70.6% which is slightly lower than previous reports. Houserová et al. (2006) found MeHg to be over 95% of total mercury in fish tissue. The low MeHg/THg proportion indicates that conditions for methylation in the sediments were unfavourable. The rate of methylation is not only a function of total mercury concentrations but also of activity of methylating bacteria such as *Methanobacterium* (Gilmour and Henry 1991; Hamasaki et al. 1995). Compared to the results from 2007, a decrease in the MeHg proportion was observed (Table 2) on sites 1 and 5.

Total mercury in sediment

As in 2007, observations showed that mercury sediment levels did not vary widely within a single year. The exception was the concentration in September samples from site 2 (0.905 mg·kg⁻¹), site 4 (0.239 mg·kg⁻¹), and site 7 (0.318 mg·kg⁻¹). While a marked increase was found on site 2, a decrease was found on sites 4 and 7. The increase could be caused by a single atypical event. Total mercury content increased along the river with the highest

Table 4. Hazard indices for a standard consumer and a member of a fishing family and maximum weekly tolerable mercury intake (kg)

Sample site	Hazard index ¹		Maximum weekly tolerable intake ²
	standard consumer	fishing family	
1 Bilovice nad Svitavou*	0.018	0.137	1.6
2 Svitava before junction*	0.010	0.078	3.0
3 Kníničky**	0.013	0.097	2.3
4 Svatka before junction**	0.021	0.160	1.7
5 Modřice**	0.012	0.090	2.7
6 Rajhradice**	0.014	0.112	2.1
7 Židlochovice**	0.014	0.112	1.8

* The Svitava River; ** the Svatka River, ¹calculation of total mercury according to Kannan et al. (1998); ²calculation of methyl mercury according to WHO (1990)

values downstream at site 6 in both 2007 and 2008, demonstrating an effect of Brno on mercury contamination in the sediment. The average THg content in sediment from 2008 is lower than in 2007 at the same sites (Table 3). Higher levels were found at sites 2 and 4 in comparison to upstream sites (1 and 3) showing the negative effect of Brno with increased pollution by waste water from households, chemical factories, and industry. Similarly to results in 2007 (Kružíková et al. 2009) the highest value was $1.15 \pm 0.18 \text{ mg}\cdot\text{kg}^{-1}$.

Although the highest values for THg in the sediment were found on site 6, the highest concentration in fish muscle was detected on site 4. One of the lowest values of THg in sediment was found at this site, indicating that the content in sediment does not correspond with the concentration of THg and MeHg in fish muscle.

All results from 2007 and 2008 were expressed as DW but Svobodová et al. (1988) have reported values of total mercury in sediment relative to organic matter. Characteristics of the analyzed sediment could impact mercury content in the muscle of fish. It is necessary to relate mercury content in the dry sediment to organic matter. Dissolved organic matter is known to promote (Weber 1993) or inhibit (Barkay et al. 1997) the formation of toxic and bioaccumulative methylmercury species.

Health hazard assessment

The hazard index (HI) for a standard consumer and a member of a fishing family were calculated for all sites according to the method set by Kannan et al. (1998) (Table 4). The calculated HI is compared to a value of 1. An HI below 1 indicates no hazard, an index ≥ 1 represents a hazard for the consumer. The average consumption of fish is used for this calculation. Recommended fish consumption is about 17 kg per capita per year. The average consumption of fish per individual worldwide is 16 kg per year, and only 11 kg per year in Europe. In the Czech Republic, the average consumption of fish is much lower than these values being 5.5 kg (for freshwater fish 1.3 kg, 4.2 kg for marine fish) (Ministry of Agriculture 2009). The average consumption of 10 kg per year was used for a fishing family (Table 4).

Hazard indices calculated for tested sites were several times lower than the hazard limit. A low HI indicated no significant health risk associated with the consumption of fish from the sites monitored in this study. The HI for both a standard consumer and a member of a fishing family was found in fish from site 4 but was still much lower than site 1. Similar results were obtained from fish sampled in 2007 (Kružíková et al. 2009).

The maximum weekly tolerable intake (Table 4) indicates the amount of fish meat that can be consumed per week without constituting a health risk. In view of the MeHg contamination, the highest quantity of fish can be consumed from sites 2 and 5, and the lowest from sites 1 and 4. Kružíková et al. (2009) identified the amount of fish that can be safely eaten from site 5 as 2.57 kg per week, a similar result to our study. Fish from the Svitava and Svatka rivers in the Czech Republic show very low HI, and their consumption poses no health risk from THg and MeHg contamination.

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References

- Caricchia AM, Minervini G, Soldati P, Chiavarini S, Ubaldi C, Morabito R 1997: GC-ECD determination of methylmercury in sediment samples using a SPB-608 capillary column after alkaline digestion. *Microchem J* **55**: 44-55
- Barkay T, Gillman M, Turner RR 1997: Effect of dissolved organic carbon and salinity on availability of mercury. *Appl Environ Microbiol* **63**: 4267-4271

- Blahová J, Havelková M, Kružíková K, Kovářová J, Haruštiaková D, Kasíková B, Hypr D, Jurčíková J, Ocelka T, Svobodová Z 2009. Fish biochemical markers as a tool for pollution assessment on the Svitava and Svatka rivers, Czech Republic. *Neuroendocrin Lett* **30** (Suppl. 1): 211-218
- Blahová, J, Havelková M, Kružíková K, Hilscherová K, Halouzka R, Modrá H, Grabic R, Halířová J, Jurčíková J, Ocelka T, Haruštiaková D, Svobodová Z 2010: Assessment of contamination of the Svitava and Svatka rivers in the Czech Republic using selected biochemical markers. *Environ Toxicol Chem* **29**: 541-549
- Čelechovská O, Svobodová Z, Žlábek V, Macharáčková B 2007. Distribution of metals in tissues of the common carp (*Cyprinus carpio*, L.). *Acta Vet Brno* **76**: S93-S100
- Commission Regulation (EC) No 1881/2006 of 19 December 2006 setting maximum levels for certain contaminants of foodstuffs
- Dušek L, Svobodová Z, Janoušková D, Vykusová B, Jarkovský J, Smid R, Pavlis P 2005: Bioaccumulation of mercury in muscle tissue of fish in the Elbe River (Czech Republic): Multispecies monitoring study 1991-1996. *Ecotox Environ Safe* **61**: 256-267
- Eisler R 2006. Mercury. Hazards to living organisms. Taylor & Francis group, New York, 312 p.
- Gilmour CC, Henry EA 1991: Mercury methylation in aquatic systems affected by acid deposition. *Environ Pollut* **71**: 131-169
- Grabic R, Jurčíková J, Tomšejová S, Ocelka T, Halířová J, Hypr D, Kodeš V 2010: Passive sampling methods for monitoring endocrine disruptors in the Svatka and Svitava rivers in the Czech Republic. *Environ Toxicol Chem* **29**: 550-555.
- Halbach S 1995: Toxicity of detrimental metal ion In: Berthon G. (Ed.): Handbook of metal-ligand interactions in biological fluids-bioinorganic medicine, Marcel Dekker, Basel, pp. 749-754
- Hamasaki T, Nagase H, Yoshioka Y, Sato T 1995: Formation, distribution, and ecotoxicity of methylmetals of tin, mercury, and arsenic in the environment. *Crit Rev Environ Sci Technol* **25**: 45-91
- Houserová P, Janák K, Kubáň P, Pavlíčková J, Kubáň V 2006: Chemical forms of mercury in aquatic ecosystems - Properties, levels, cycle and determination. *Chem Listy* **100**: 862-876
- Kannan K, Smith RG, Lee RF, Vindom HL, Heitmuller PT, Macauley JM, Summers JK 1998: Distribution of total mercury and methylmercury in water, sediment and fish from south Florida estuaries. *Arch Environ Contam Toxicol* **34**: 109-118
- Kružíková K, Randák T, Kenšová R, Kroupová H, Leontovyčová D, Svobodová Z 2008a: Mercury and methylmercury concentration in muscle tissue of fish caught in major rivers of the Czech Republic. *Acta Vet Brno* **77**: 637-643
- Kružíková K, Svobodová Z, Valentová O, Randák T, Velišek J 2008b: Mercury and methylmercury in muscle tissue of chub from the Elbe river main tributaries. *Czech J Food Sci* **26**: 65-70
- Kružíková K, Blahová J, Kenšová R, Jurčíková J, Hypr D, Svobodová Z 2009: Mercury and methylmercury content in chub from the Svitava and Svatka rivers at agglomeration Brno. *Czech J Food Sci* **27**: 470-476
- Maršálek P, Svobodová Z 2006: Rapid determination of methylmercury in fish tissues. *Czech J Food Sci* **24**: 138-142
- Maršálek P, Svobodová Z, Randák T 2006: Total mercury and methylmercury contamination in fish from various sites along the Elbe River. *Acta Vet Brno* **75**: 579-585
- Morel FMM, Kraepiel AML, Amyot M 1998: The chemical cycle and bioaccumulation of mercury. *Annu Rev Ecol Syst* **29**: 543-566
- Ministry of Agriculture 2009: Situation and outlook report fish (in Czech), Dec. 2009
- Svobodová Z, Hejtmánek M, Příkryl I, Kocová A 1988: The content of total mercury in the individual components of the ecosystem of the Želivka water reservoir I. Water, sediments. In Czech: Obsah celkové rtuti v jednotlivých složkách ekosystému vodárenské nádrže Želivka I. Voda, sedimenty. *Buletin VÚRH Vodnany* **24**: 26-33
- Swain EB, Engstrom DR, Brigham ME, Henning TA, Brezonik PL 1992: Increasing rates of atmospheric mercury deposition in midcontinental North America. *Science* **257**: 784-787
- Weber JH, 1993: Review of possible paths for abiotic methylation of mercury (II) in the aquatic environment. *Chemosphere* **26**: 2063-2077
- Weiner JG, Spry DP 1996: Toxicological significance of mercury in freshwater fish. In: Beyer WM, Heinz GH Redmon-Norwood AW (Eds): Environmental contaminants in wildlife. Interpreting tissue concentrations. Ewis, Boca Raton f.l. USA, pp. 297-339
- WHO (World Health Organization) 1990: Methylmercury. In: Environmental Health Criteria. World Health Organisation (WHO). Geneva. 145. www.who.org/mercury. Cited 30 Jan 2009
- Žlábek V, Svobodová Z, Randák T, Valentová O 2005: Mercury content in the muscle of fish from the Elbe River and its tributaries. *Czech J Anim Sci* **50**: 528-534